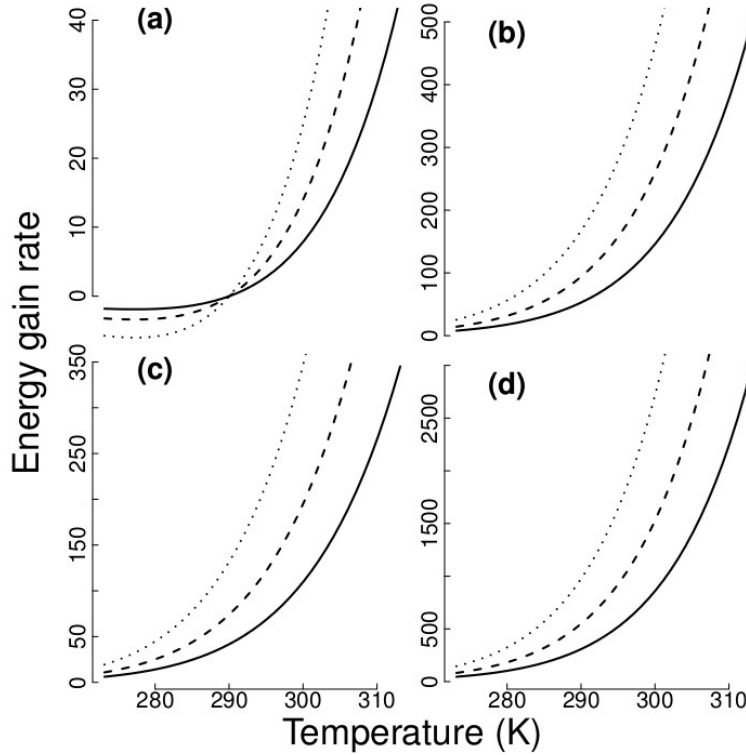


## Supporting Information for "Ecological emergence of thermal clines in body size"

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### 1. Temperature-dependent allometric scaling of net energy gain



**Figure S1.** Mass-specific energy gain rate (defined as  $I-M$  in Eq. S1) as a function of temperature in the range 0-40°C (273-313 K) for differently sized vertebrate ectotherms (a, c; dotted line:  $m_c = 1\text{g}$ , dashed line:  $m_c = 10\text{g}$ , solid line:  $m_c = 100\text{g}$ ) and multicellular invertebrates (b, d; dotted line:  $m_c = 0.01\text{g}$ , dashed line:  $m_c = 0.1\text{g}$ , solid line:  $m_c = 1\text{g}$ ) in which ingestion is either ecologically-limited ( $f_i=0.2$ , a, b) or physiologically-limited ( $f_i=0.9$ , c; d). Other parameters are as follows: vertebrate ectotherms:  $a_I=6.4$ ,  $a_M=2.3$ ,  $E_I=0.67$ ,  $E_M=0.43$ ; multicellular invertebrates:  $a_I=9.7$ ,  $a_M=0.51$ ,  $E_I=0.77$ ,  $E_M=0.79$ .

Following Vasseur & McCann (2005) and references therein, we write the mass-specific biological rate functions including body mass and temperature scaling as:

$$\begin{cases} I = f_I a_I(T_0) m_c^{-0.25} e^{E_I(T-T_0)/kTT_0} \\ M = a_M(T_0) m_c^{-0.25} e^{E_M(T-T_0)/kTT_0} \end{cases} \quad \text{Eq. S1,}$$

where  $I$  is energy intake rate,  $M$  is the rate of energy loss to metabolism,  $m_c$  is consumer body mass,  $a_i(T_0)$  are empirically-derived intercepts of the allometric relationships (in  $\text{kg} (\text{kg year})^{-1} \text{kg}^{0.25}$ ) which represent the maximum sustainable rates (physiological maxima) measured at temperature  $T_0$ ,  $T$  is absolute temperature (in K),  $k$  is Boltzmann's constant ( $8.618 \times 10^{-5} \text{ eV K}^{-1}$ ),  $E_a$  is activation energy (in eV) of the reaction, and  $f_i$  represents the realized fraction (in the wild) of the physiological ingestion

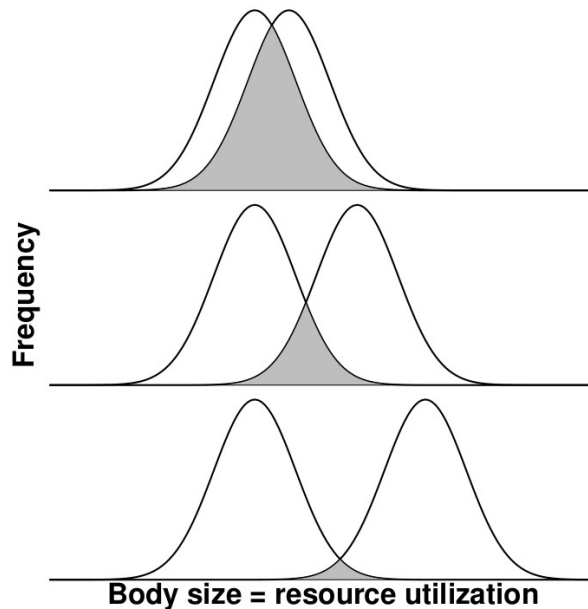
maxima that can be attained under ideal conditions. When species are “ecologically limited” for ingestion  $f_I \ll 1$ , while  $f_I \approx 1$  when species are “physiologically-limited” for ingestion (Vasseur & McCann, 2005). In Fig. S1, we posit mass-specific energy gain rate  $G = I - M$  (Ohlberger *et al.*, 2011), and represent  $G$  as a function of  $T$  for differently-sized vertebrate ectotherms (Figs. S1a and c) and multicellular invertebrates (Figs. S1b and d) under either ecological ( $f_I=0.2$ , Figs. S1a and b) or physiological ( $f_I=0.9$ , Figs. S1c and d) limitation for ingestion (other parameters are as follows: vertebrate ectotherms:  $a_I=6.4$ ,  $a_M=2.3$ ,  $E_I=0.67$ ,  $E_M=0.43$ ; multicellular invertebrates:  $a_I=9.7$ ,  $a_M=0.51$ ,  $E_I=0.77$ ,  $E_M=0.79$ ). In all cases,  $G$  increases with temperature, consistent with the general finding that growth efficiency increases at higher temperatures (Angilletta & Dunham, 2003). However, the thermal sensitivity of  $G$  is higher in smaller organisms, so that warming generates a competitive asymmetry in favor of smaller organisms. Note that under “ecological limitation” and low temperatures, large vertebrates are more competitive than small vertebrates as they starve less fast (Fig. S1a). Interestingly, similar results were derived independently by Ohlberger *et al.* (2011) using more sophisticated data on perch *Perca fluviatilis* physiological rates, which incorporated a more realistic, hump-shaped relationship between energy gain rate and temperature (i.e., temperature optimum). This suggests that, although Arrhenius kinetics do not account for the hump-shaped nature of physiological and ecological rates (Knies & Kingsolver, 2010), they still provide fairly good approximations (Dell *et al.*, 2011).

## 2. Log-normal distributions for size-dependent niches

Hutchinson (Hutchinson, 1957) defined the ecological niche as the range of environmental conditions that allow a population growth rate to be zero or positive. The niche can be thought of as a  $n$  dimensional volume with its  $n$  axes represented by requisite resources. MacArthur & Levins (MacArthur & Levins, 1967) extend the theory by considering niche axes as “resource utilization” axes (Schoener, 2009). Instead of describing population growth variation in response to environmental variation, this resource-utilization niche describes the frequency distribution of resource use along each niche axis and directly quantifies the intensity of competition from the overlap of resource utilization niches (MacArthur & Levins, 1967; MacArthur, 1972) (Figure S2).

However, a major problem of the hypervolume approach is that its entire dimensionality generally extends beyond the technical and practical abilities of any one researcher. One way to circumvent this dimensionality problem is to collapse all resource axes onto one single, synthetic resource axis. Particularly relevant to this unidimensional approach is the use of a *consumer body size* axis (Figure S2). Indeed, body size correlates with prey size and type, ingestion rate, energy requirements, home ranges, and encounter rates (Peters, 1983; Jetz *et al.*, 2004; Brown *et al.*, 2004; Woodward *et al.*, 2005; Brose, 2010; Lang *et al.*, 2012). Hence, the strength of competition is proportional to the overlap of body size distributions between competitors (MacArthur & Levins, 1967; MacArthur, 1972). On macroevolutionary time scales, this size-dependent competition is predicted to result in character displacement and in a more or less regular spacing of potential competitors along a body size gradient; and in fact a number of examples support this prediction (Hutchinson, 1959; Pyke, 1982; Dayan *et al.*, 1989; Hermoyian *et al.*, 2002). This size-dependent, unidimensional approach to resource utilization niche can also be used to define the niche of predators for a focal species. Meta-analyses show that the average body mass of a predator is 100 times the body mass of its prey (Brose *et al.*, 2006), and the predator's niche may thus correspond to the niche of competitors, right-translated by 100 units along a body mass axis (or by 4.6 units along a body length axis since  $\text{mass} \approx \text{length}^3$ ). We apply this framework to compute interactions strengths experienced by individual fish in communities. We

considered competition to be maximal between two individual competitors whose body size ratio is equal to 1 (MacArthur & Levins, 1967; MacArthur, 1972), and predation to be maximal when the predator/prey length ratio is equal to 4.6 (Brose *et al.*, 2006).



**Figure S2.** Multidimensional resource utilization niches projected onto a single consumer body size axis. From top to bottom, the decreasing overlap gradient between consumer size distributions determines a decreasing gradient of competition strength.

Choice of the shape of the resource utilization niche around this maximum also requires some discussions. The theory of resource utilization niches was developed assuming normal (Gaussian) utilization curves (Figure S2, (MacArthur, 1972)). However, several authors stress that normal utilization niches represent a case of limited generality (Roughgarden, 1974; Wilson, 1975; Abrams *et al.*, 2008; Pigolotti *et al.*, 2010), and there seems to be no general rule for the shape of the utilization distribution except that its variance should increase with its mean because prey-size range (niche breadth) increases with predator size (Wilson, 1975; Woodward *et al.*, 2005). Here, we used a lognormal resource utilization distribution because (i) body sizes are often log-normally distributed (such as in our dataset), (ii) the variance of the log-normal distribution increases with its mean, accounting for the fact that niche breadth increases with body size, and (iii) because of its heavy right tail the lognormal distribution accounts for the fact that, within a niche, larger consumers have higher consumption rates (Brown *et al.*, 2004) and are thus likely to be stronger interactors than smaller consumers.

**3. Table S1.** Fish species included in our analysis. Species-specific number of observations, trophic guild, piscivory probability (based on diet data from Fishbase <http://www.fishbase.org/>) and mean body length in our dataset (from a log-normal distribution). Generalists: macroinvertebrates and fish; Insectivores: insects; Invertivores: insects, mollusks, and crustaceans; Omnivores: invertebrates and plants; Piscivores: fish.

Species latin name	Species common name	N	Trophic guild	Piscivory probability $\phi$	Mean body length (mm)	code
<i>Abramis brama</i>	Freshwater bream	23,784	Generalist	0.5	173.8	BRE
<i>Alburnoides bipunctatus</i>	Schneider	75,138	Omnivore	0	73.5	SPI
<i>Alburnus alburnus</i>	Bleak	128,108	Insectivore	0	73.1	ABL
<i>Ameiurus melas</i>	Black bullhead	14,975	Generalist	1	135.3	PCH
<i>Anguilla anguilla</i>	Eel	180,444	Generalist	1	359.3	ANG
<i>Aspius aspius</i>	Asp	383	Piscivore	1	106.8	ASP
<i>Barbus barbus</i>	Barbel	125,293	Generalist	1	164.2	BAF
<i>Barbus meridionalis</i>	Mediterranean barbel	23,796	Omnivore	0	115.5	BAM
<i>Blicca bjoerkna</i>	White bream	33,532	Omnivore	0	110.6	BRB
<i>Carassius sp.</i>	Crucian carp	5,924	Omnivore	0	173.4	CAS
<i>Chondrostoma nasus</i>	Common nase	29,862	Omnivore	0	229.0	HOT
<i>Chondrostoma toxostoma</i>	French nase	17,469	Omnivore	0	134.1	TOX
<i>Cobitis taenia</i>	Spined loach	4,699	Omnivore	0	76.4	LOR
<i>Cottus sp.</i>	Sculpin	269,207	Generalist	1	64.3	CHA
<i>Cottus petiti</i>	Lez sculpin	1,232	Omnivore	0	33.4	CHP
<i>Cyprinus carpio</i>	Common carp	5,804	Generalist	0.5	286.3	CCO
<i>Esox lucius</i>	Pike	23,061	Piscivore	1	271.9	BRO
<i>Gambusia holbrooki</i>	Mosquitofish	1,603	Insectivore	0	29.6	GAM
<i>Gasterosteus aculeatus</i>	Threespined stickleback	18,828	Invertivore	0	41.6	EPI
<i>Gobio sp.</i>	Gudgeon	390,104	Omnivore	0	88.8	GOU
<i>Gymnocephalus cernuus</i>	Ruffe	9,267	Generalist	1	93.9	GRE
<i>Lampetra fluviatilis</i>	River lamprey	86	Piscivore	1	122.9	LPR
<i>Lampetra planeri</i>	European brook lamprey	45,423	Omnivore	0	119.8	LPP
<i>Lepomis gibbosus</i>	Pumpkinseed	45,725	Generalist	1	80.0	PES
<i>Leucaspius delineatus</i>	Sunbleak	2,181	Insectivore	0	45.7	ABH
<i>Leuciscus burdigalensis</i>	Rostrum dace	418	Insectivore	0	159.9	VAR
<i>Leuciscus cephalus</i>	Chub	419,627	Generalist	1	160.0	CHE
<i>Leuciscus idus</i>	Orfe	208	Generalist	0.5	98.1	IDE
<i>Leuciscus leuciscus</i>	Common dace	78,193	Insectivore	0	134.9	VAN
<i>Lota lota</i>	Burbot	4,963	Piscivore	1	244.3	LOT
<i>Micropterus salmoides</i>	Largemouth bass	1,459	Piscivore	1	115.7	BBG
<i>Misgurnus fossilis</i>	Weatherfish	1,961	Omnivore	0	140.8	LOE
<i>Barbatula barbatula</i>	Stone loach	281,339	Omnivore	0	69.5	LOF
<i>Onchorhynchus mykiss</i>	Rainbow trout	3,813	Generalist	1	217.1	TAC
<i>Pachychilon pictus</i>	Albanian Roach	1,815	Generalist	0.5	91.9	PAP
<i>Perca fluviatilis</i>	Perch	87,796	Piscivore	1	124.8	PER
<i>Phoxinus phoxinus</i>	Eurasian minnow	367,104	Omnivore	0	55.4	VAI
<i>Pseudorasbora parva</i>	Topmouth gudgeon	9,323	Generalist	0.5	58.4	PSR
<i>Pungitius pungitius</i>	Ninespined stickleback	16,234	Invertivore	0	37.7	EPT
<i>Rhodeus amarus</i>	Bitterling	39,369	Insectivore	0	46.7	BOU
<i>Rutilus rutilus</i>	Roach	356,333	Omnivore	0	120.9	GAR
<i>Salaria fluviatilis</i>	Freshwater blenny	1,217	Omnivore	0	63.3	BLE
<i>Salmo salar</i>	Atlantic salmon	56,089	Generalist	1	103.0	SAT
<i>Salmo trutta</i>	Brown trout	846,862	Generalist	1	144.2	TRF
<i>Salvelinus fontinalis</i>	Brook trout	412	Generalist	1	202.8	SDF
<i>Scardinius erythrophthalmus</i>	Rudd	16,527	Omnivore	0	108.5	ROT
<i>Silurus glanis</i>	Wels catfish	6,149	Generalist	1	285.3	SIL
<i>Sander lucioperca</i>	Pikeperch	3,640	Piscivore	1	222.3	SAN
<i>Telestes souffia</i>	Vairone	58,847	Omnivore	0	101.8	BLN
<i>Thymallus thymallus</i>	Grayling	9,419	Generalist	0.5	201.3	OBR
<i>Tinca tinca</i>	Tench	16,530	Invertivore	0	160.2	TAN
<i>Zingel asper</i>	Rhone streber	51	Omnivore	0	121.6	APR

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